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Identifying the Sources and Sinks of CDOM/FDOM across the Mauritanian Shelf and Their Potential Role in the Decomposition of Superoxide ($O_2^-$)

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Superoxide ($O_2^-$) is a short lived reactive oxygen species (ROS) formed in seawater by photochemical or biological sources, it is important in the redox cycling of trace elements and organic matter in the ocean. The photoproduction of $O_2^-$ is now thought to involve reactions between $O_2$ and reactive reducing (radical) intermediates formed from dissolved organic matter (DOM) via intramolecular reactions between excited singlet state donors and ground-state acceptors (Zhang et al., 2012). In seawater the main pathways identified for the decomposition of $O_2^-$ into $H_2O_2$ and $O_2$ involve reactions with Cu, Mn, and DOM. In productive regions of the ocean, the reaction between DOM and $O_2^-$ can be a significant sink for $O_2^-$. Thus, DOM is a key component of both the formation and decomposition of $O_2^-$ and formation of $H_2O_2$. In the present work we examined the relationships between $O_2^-$ decay rates and parameters associated with chromophoric dissolved organic matter (CDOM) and fluorescent dissolved organic matter (FDOM) by using the thermal $O_2^-$ source SOTS-1. Filtered samples (0.2  $\mu$m) were run both in the presence, and absence, of the metal chelator diethylenetriaminepentaacetic acid (DTPA) to determine the contribution from DOM. Samples were collected along a transect across the continental shelf of the Mauritanian continental shelf during a period of upwelling. In this region we found that reactions with DOM, are a significant sink for $O_2^-$ in the Mauritanian Upwelling, constituting on average 58 ± 13% of the $O_2^-$ loss rates. Superoxide reactivity with organic matter showed no clear correlation with bulk CDOM or FDOM properties (as assessed by PARAFAC analysis) suggesting that future work should concentrate at the functional group level to clearly elucidate which molecular species are involved as bulk properties represent a wide spread of chemical moieties with different $O_2^-$ reactivities. Analysis of FDOM parameters indicates that many of the markers used previously for terrestrial sources of DOM and FDOM are called into question as marine sources exist. In particular recent work (Rico et al., 2013) indicates that algal species may also produce syringic, vanillic, and cinnamic acids, which had previously been ascribed solely to terrestrial vegetation.

Keywords: reactive oxygen species, parafac, colored dissolved organic matter (CDOM), Atlantic Ocean, excitation emission matrix fluorescence, fluorescence dissolved organic matter (FDOM), superoxide dismutase, hydrogen peroxide
INTRODUCTION

Reactive Oxygen Species in Seawater—Superoxide ($O_2^-$) and Peroxide ($H_2O_2$)

Superoxide ($O_2^-$) is an important transient reactive oxygen species (ROS) in the ocean formed as a reactive intermediate in photosynthesis and respiration and with this the conversion of oxygen ($O_2$) into water and vice versa. $O_2^-$ is a highly reactive and the short-lived radical anion can be produced both via photochemical (Micinski et al., 1993) and biological processes in the ocean (Diaz et al., 2013; Roe et al., 2016). $O_2^-$ and $H_2O_2$ are directly involved in degradation of organic pollutants and photobleaching of CDOM (Scully et al., 2003; Chen et al., 2009), cause oxidative stress in aquatic organisms and alter the redox cycling of trace metals like Fe, Cu, and Mn (Moffett and Zika, 1987; Wuttig et al., 2013b). In earlier work, where we examined the decomposition rate of $O_2^-$ throughout the water column in the Eastern Tropical North Atlantic (ETNA) Ocean, we found that in the surface ocean, which is in this area strongly impacted by Saharan aerosols and coastal sediment resuspension, the main decay pathways for superoxide (Figure 1) were reactions with Mn(II) and DOM (Wuttig et al., 2013a,b).

Chromophoric Dissolved Organic Matter (CDOM)

DOM is a complex mix of organic molecules and is poorly described in terms of its composition. CDOM is the proportion of DOM that absorbs light and this can be characterized by its absorbance and fluorescence properties (Coble, 2007). CDOM is a ubiquitous component of the open ocean dissolved matter pool, and is important because of its influence on the optical properties of the water column, its role in photochemistry and photobiology, and its utility as a tracer of deep ocean biogeochemical processes and circulation. The general distribution of CDOM in the global ocean is controlled by a balance between production and photolysis, with vertical circulation playing an important role in the transport of CDOM to and from intermediate water masses. Fluctuations in the abundance of CDOM in the global surface ocean have been observed, indicating a potentially important role for CDOM in ocean-climate connections because of its impact on photochemistry and photobiology (Nelson and Siegel, 2013).

Pioneering work by Coble (1996) showed that the Excitation Emission Matrix (EEM) measurements of CDOM fluorescence (often referred to as FDOM) can generally be divided into two categories—humic-type or protein/amino acid-type fluorescence. Furthermore, Coble (1996) defined 5 major fluorescence regions as per the excitation/emission spectra as follows: Humics—peak A ($\lambda_{ex}/\lambda_{em}$ ∼260/380–460 nm), Peak C ($\lambda_{ex}/\lambda_{em}$ ∼350/420–480 nm), peak M ($\lambda_{ex}/\lambda_{em}$ ∼312/380–420 nm); Proteins—peak B ($\lambda_{ex}/\lambda_{em}$ ∼275/310 nm) and peak T ($\lambda_{ex}/\lambda_{em}$ ∼275/340 nm). More recently EEM has been combined with parallel factor (PARAFAC) data analysis (Stedmon and Bro, 2008) to independently determine multiple components of the CDOM pool—many, but not all, of which are related.

![Figure 1](image_url)  
**Figure 1:** Schematic of the different decay pathways for $O_2^-$ decay in the ocean. $O_2^-$ is biologically and photo produced. Modified from Wuttig et al. (2013a).
to the peaks found in the original Coble analysis. Humic-like fluorescence (FDOM$_{H}$; typically $\gamma_{ex}/\lambda_{em} \sim 320/420$ nm) has been observed in a wide range of marine environments correlates in general well with nutrients (NO$_3^-$, PO$_4^{3-}$) and apparent oxygen utilization (AOU) in different water masses (Hayase and Shinozuka, 1995; Kuma et al., 1998; Yamashita et al., 2007; Yamashita and Tanoue, 2008). These correlations suggest that some of the components that make up FDOM are formed by the remineralization of settling organic particles and are destroyed or modified by irradiation. However, marine humic substances are composed of a large fraction of the uncharacterized DOM pool in the ocean (Zafiriou et al., 1984) and the relative contribution of these complex substances to seawater fluorescence is still unclear.

**Role of CDOM in the Production and Decomposition of ROS in Seawater**

Our understanding of how ROS species are generated in seawater by CDOM absorption of sunlight in the euphotic zone of the ocean has advanced substantially in recent years. In particular the paradigm that existed until recently, that excited triplet states of CDOM reacted with O$_2$ to form O$_2^-$ and carbocations (O’Sullivan et al., 2005) has been replaced with a new mechanism in which a low-efficiency intramolecular electron transfer from an excited singlet donor (e.g., substituted phenol) to a ground-state acceptor (e.g., quinone), that produces a radical species that reacts with O$_2$ to form O$_3^-$ and subsequently H$_2$O$_2$ (Zhang et al., 2012; Sharpless and Blough, 2014). Recent work by Powers and Miller (2014) estimated using satellite climatologies, using average apparent quantum yield (AQY) spectrum determined from laboratory irradiations, found that daily H$_2$O$_2$ photoproduction rates (averaged over an annual cycle) were highest in equatorial regions and lowest at the poles (range: 0.07–93.2; average: 40.4; median: 39.5 nM per day). The same group in a further paper, Powers et al. (2015), also reevaluated the photoreactivity of refractory DOC by investigating the photochemical production H$_2$O$_2$ and O$_2^-$, using controlled irradiations at sea and in the laboratory. They found that in the open ocean, a large fraction of photoproduced O$_2^-$ does not lead to H$_2$O$_2$, which means, that the relationship between these two ROS involve complex pathways. In particular, the apparent stoichiometry of formation was found to be closer to 4:1 (Powers et al., 2015) instead of the 2:1 which would be expected solely from dismutation. This may in part be explained by the photo-generation of oxidized CDOM species, which can act as an electron acceptor, and react with O$_2^-$ to form O$_2$ (Heller et al., 2010). This in turn may be a significant sink for refractory DOC as it circulates through the surface ocean (Mopper et al., 1991; Stubbins et al., 2012).

It is well-known that photo-oxidation of proteins such as Tryptophan produces O$_2^-$ and subsequently H$_2$O$_2$ (McCormick and Thomason, 1978), reactions with other photo-produced reactive oxygen species (ROS) (e.g., O$_2$ and OH) may also be important pathways for the destruction of proteins, and hence the loss of protein-like fluorescence, in the ocean (Boreen et al., 2008). Similarly the loss of FDOM in surface waters is often ascribed to photo-bleaching which may be due to reactions with O$_2^-$ (Omori et al., 2011). Thus, quantitative information on the production, sinks, and concentrations of O$_2^-$ and H$_2$O$_2$ in the open ocean is fundamental to fully understand their role in global biogeochemical cycles (Powers and Miller, 2014).

**Study Region—Mauritanian Upwelling**

In the present work we focus on the Mauritanian upwelling system which stretches from the Iberian Peninsula to about 10°N along the Northwest African coast. The Mauritanian upwelling is one of the main Eastern Boundary Upwelling Systems (EBUS), where nutrient rich waters are upwelled by the trade winds to fuel one of the most biologically productive regions in the global ocean (Messié and Chavez, 2015). Due to changes in wind forcing the coastal upwelling off Mauritania exhibits a pronounced seasonal cycle and the Mauritanian upwelling is the most productive branch of the Canary Current upwelling system (Tanhua and Liu, 2015). Upwelling between 20° N and 25° N is persistent throughout the year (Schaafstall et al., 2010). In contrast, upwelling north and south of this area is strongly seasonal due to wind forcing associated with the migration of the ITCZ (Tomczak and Godfrey, 1994). Vertical mixing induced by bottom turbulence is also an important transport process for supplying nutrients to the euphotic zone (Schaafstall et al., 2010). Primary production is high year-round (80–200 mmol m$^{-2}$ d$^{-1}$ of C) and elevated beyond the shelf break (Huntsman and Barber, 1977).

The seasonality in the upwelling strength also impacts phytoplankton dynamics of the Senegal-Mauritanian upwelling region (Farikou et al., 2015), as a seasonal cycle is observed beginning with the onset of the upwelling (December–February), mainly nanoeukaryote type phytoplankton are found in the coastal area; while in April–May, the period corresponding to the maximum chlorophyll a concentration, the nanoeukaryotes were replaced by diatoms.

The Oxygen Minimum Zone (OMZ) in the ETNA appears to be undergoing a significant water column deoxygenation of 0.5 μmol kg$^{-1}$ yr$^{-1}$ (Stramma et al., 2009).

**Aims and Objectives of the Present Study**

Our main aim in this study was to assess the role of DOM in the decomposition of O$_2^-$ along a transect across the continental shelf in the Mauritanian upwelling region by comparing O$_2^-$ decay rates to bulk CDOM and FDOM properties. A further objective was to identify CDOM or FDOM parameters that may indicate what type of organic material is responsible for the production or decomposition of O$_2^-$ in the ocean. The final objective was to examine the influence of primary production, photobleaching, and microbial activity on CDOM and FDOM properties in an upwelling region devoid of riverine input.

**MATERIALS AND METHODS**

In order to prevent trace metal contamination all laboratory work was performed at sea in a trace metal clean chemistry laboratory under ISO Class 5 conditions using a specially designed containerized clean room (Clean Modules UK). All
chemicals that were used in this study were of ultrapure grade unless noted. Ultrapure (UP) water (resistivity > 18.2 MΩ cm⁻¹) was obtained in the laboratory and in the ship going clean container via a Millipore Synergy 185 system that was fed by an Elix-3 (Millipore) reverse osmosis system connected to the mains supply. Pipettes (Finnpipette) were calibrated monthly and trace metal clean pipette tips (Rainin RT-250, RT-1000, and Finntip 10) were used as supplied. An inoLab pH 720 (WTW) was used to determine pH values on the NBS scale (pH_{NBS}). All plasticware and bottles (low density high polyethylene (LDPE) and Polytetrafluoroethylene (PTFE)) were extensively cleaned according to the GEOTRACES trace metal clean protocols (Cutter et al., 2010).

Seawater Sampling
Seawater samples were collected using Go-Flo sampling bottles deployed on a Kevlar line at 5 stations (Figure 2) occupied during the RV Maria S. Merian expedition MSM17/4 from March 13 to
April 5 2011 (boreal spring, upwelling season). The E-W transect along 18°N across the shelf/slope at the Mauritanian upwelling region covered a distance of 50 km where water depth increases from ~50 to ~1100 m (please See Table S1 for the details on all of the stations occupied and samples taken).

**Hydrography and Nutrients**

Hydrographical data was obtained using a Seabird Conductivity-Temperature-Depth (CTD) rosette system. This system consisted of a SBE911plus CTD system in combination with a carousel water sampler SBE32 with 24 12-L bottles. The CTD system was equipped with a CT sensor pair, two O2 sensors (SBE43 SN871 and SBE43 SN950 -calibrated by comparison to Winkler O2 titrations), a turbidity sensor and a chlorophyll sensor (chlorophyll a fluorescence—calibrated according to the manufacturer’s protocols). CDOM and FDOM samples were obtained from Go-Flo bottles and Niskins on the CTD system. Samples for dissolved macro nutrients were subsampled from the Go-Flo bottles and either analyzed onboard for low level (nM) concentrations of nitrate (NO3−) and ammonium (NH4+) using photometric (Grasshoff et al., 1999) or fluorometric method (Holmes et al., 1999), respectively. Samples for silica (H2SiO4), phosphate (PO43−) and nitrate (NO3−) were frozen, transported back to Germany for analysis in the Nutrient laboratory of the MPI-Bremen (Grasshoff et al., 1999).

**Satellite Data**

Satellite chlorophyll-a data was obtained from OBPG MODIS-Aqua Monthly Global 9-km Products via GIOVANNI (http://giovanni.gsfc.nasa.gov/giovanni/) using the Ocean Color Time-Series Online Visualization and Analysis platform. Analyses and visualizations used in this paper were produced with the Giovanni online data system, developed and maintained by the NASA Goddard Earth Sciences (GES) Data and Information Services Center (DISC). All satellite images were finally displayed as postscript images using the Generic Mapping Tools (GMT) software (Wessel and Smith, 1998).

**CDOM Absorbance and Fluorescence**

**CDOM Absorbance**

CDOM absorbance measurements were performed using a liquid waveguide capillary cell (LWCC) (LWCC-2100 World Precision Instruments, Sarasota, FL, USA) and an Ocean Optics USB4000 UV-VIS spectrophotometer in conjunction with an Ocean Optics DT-MINI-2-GS light source. Samples were filtered through an 0.2 μm syringe filter (Sarstedt) using a 10 mL Teflon syringe (Savillex), the first ~10 mL were discarded and the absorbance of the afterwards filtered solution then measured by direct injection into the LWCC. Absorbance measurements were made relative to UP and corrected for the refractive index of seawater based on the procedure outlined in Nelson et al. (2007). The resulting dimensionless optical density spectra were converted to absorption coefficient (m−1): aCDOM(λ) = 2.303 Aλ/λ, where 2.303 converts decadal logarithmic absorbance to base e, and λ is the effective optical pathlength of the waveguide (here 100.3 ± 0.5 cm as determined by the manufacturer). In the present work we measured CDOM absorbance over the wavelength (λ) range 280–800 nm.

The spectral slope parameter (Helms et al., 2008), S, was calculated over a range of 275–295 nm (S275−295) and 350–400 nm (S350–400) using a non-linear least squares fitting procedure in Matlab. S275−295 is commonly used as a proxy for molecular weight with increasing values indicating decreasing molecular weight and aromaticity (Helms et al., 2008). Similarly the E2:E3 ratio, also used to track changes in the relative size of CDOM, was calculated as the ratio of CDOM absorption at 250 to 365 nm (De Haan and De Boer, 1987).

**CDOM Fluorescence—Excitation Emission Matrix**

Samples for CDOM fluorescence measurements were syringe filtered through 0.2 μm filters (Sarstedt) as described above for the absorbance measurements. Humic-type fluorescence measurements were performed with a Cary Eclipse Fluorometer using a 1 cm quartz cell. Measurements of FDOMH (Tani et al., 2003) were performed by analysis of samples using excitation at 320 nm and emission at 420 nm (10 nm slit widths). Each sample was also analyzed as Excitation Emission Matrix (EEM) on the Cary Eclipse Fluorometer using the same 1 cm quartz cell as for the FDOMH measurements. For the EEM analysis, excitation wavelengths were scanned (12000 nm/min) from 250 to 500 nm (5 nm slit width and 5 nm increments) and emission wavelengths (5 nm slit width and 5 nm increments) from 280 to 600 nm, the photon multiplier tube (PMT) voltage was set at 700 V (maximum) and the response time 0.08 s. Day to day variation in the instrument was monitored by daily measurements of the Raman scattering of UP water (excitation 350 nm; Stedmon et al., 2003; Heller et al., 2013) and the use of a standard of quinine sulfate (1 ppm in 0.05 N H2SO4) which was also diluted to form a calibration series for quinine fluorescence (QSU) and run daily (Mopper and Schultz, 1993).

Post-processing of the complete EEM data set (Go-Flo and Niskin bottles) was performed according to accepted protocols (Lawaetz and Stedmon, 2009; Murphy et al., 2010) in the following sequence: (i) correction of instrument bias using the correction files provided by the manufacturer, (ii) subtraction of the EEM of UP water, and finally, (iii) the fluorescence intensity was normalized to the area under the UP water Raman peak (excitation 350 nm) run with each sample batch. Note as the sample absorbance was low, no correction was made for the internal absorption of the samples.

**PARAFAC Analysis of 3D Excitation Emission Matrix (EEM)**

The normalized EEMs were analyzed by PARAFAC in MATLAB under application of the DOMfluor toolbox (Stedmon and Bro, 2008) using models constrained to non-negative values. Outlier identification was performed using the outlier test function provided with the DOMfluor toolbox. No samples with extreme leverage were found, indicating no extreme, and potentially outlying, EEMs in the dataset. Determination of the most suitable number of components was achieved by the split-half analysis and random initialization where by both halves were successfully
validating. No systematic residual was found in the modeled EEMs.

**Determination of $\text{H}_2\text{O}_2$**

Samples for $\text{H}_2\text{O}_2$ were analyzed directly using a flow injection chemiluminescence (FIA-CL) reagent injection method (Yuan and Shiller, 1999) as described previously (Croot et al., 2004). Samples were analyzed using 5 replicates; typical precision was 2–3% through the concentration range 0.5–100 nM, the detection limit (3 s) is typically 0.2 nM.

**Determination of $\text{O}_2^-$ Decay Rates**

**Experimental Design**

In the present study we employed the thermal $\text{O}_2^-$ source SOTS-1 [di(4-carboxybenzyl) hyponitrite—molecular weight 330.3 g mol$^{-1}$; Ingold et al., 1997] as described by us previously (Heller and Croot, 2010a). SOTS-1 has some advantages over KO$_2$ and other currently used methods which generate $\text{O}_2^-$ at $\mu$M concentrations as it produces $\text{O}_2^-$ slowly and continuously over a longer duration to be able to mimic the *in vivo* situation and additionally there is little or no $\text{H}_2\text{O}_2$ formed upon introduction to the sample. It also avoids the problem of adding the chelator DTPA, in order to complex metals, prior to irradiating ketone solutions to produce $\text{O}_2^-$ (Mcdowell et al., 1983), as photodegradation products of DTPA will be formed and cause problems with calibration and speciation analysis (Heller and Croot, 2011).

The decomposition rate of SOTS-1 is well-described in seawater and was shown to follow a first order decay with a 40 mol% yield of $\text{O}_2^-$ (Ingold et al., 1997).

500 $\mu$g aliquots of SOTS-1 were used as received (Cayman Chemicals) and stored at $-80^\circ$C until use. Immediately prior to the start of any experiment a primary stock of SOTS-1 was prepared by the dissolution of the 500 $\mu$g SOTS-1 aliquots in DMSO (Fluka, puriss p.a.=99.9%) before further dilution in seawater. Final starting concentrations for SOTS-1 ([SOTS-1]$_0$) at the beginning of each experiment in this study were between 0.86 and 1.78 $\mu$M (Table S2). Experiments were performed in Teflon bottles (Nalgene) which were either left unamended or had DTPA, Cu (0.79, 1.58 nM), Mn (1.00, 2.00 nM), or Fe (0.90, 1.79 nM) added. All samples were equilibrated for at least 12 h before the experiment was initiated by the addition of a specific amount of SOTS-1 from the primary standard to a known volume of seawater. All reagents and samples were kept at a constant temperature (21.5 $\pm$ 0.2$^\circ$C) throughout the course of the experiment in the temperature controlled class 100 clean laboratory. Only the unamended and DTPA results are reported in the present work, the results from the trace metal additions will be reported elsewhere.

**Determination of $\text{O}_2^-$ Concentrations using MCLA Chemiluminescence**

The most widely used approach to measuring $\text{O}_2^-$ in seawater (Rose et al., 2008; Heller and Croot, 2010a) is via the use of the chemiluminescence probe Cypridina luciferin analog [[2-methyl-6-(4-methoxyphenyl)-3,7-dihydroimidazo[1,2-a]pyrazin-3-one, HCl]] (MCLA) (Fluka) (Nakano et al., 1986). In the present work we used the same method as we had used previously (Heller and Croot, 2010a,c). A brief description follows; MCLA was used as received, a primary 1 mM MCLA standard was prepared by dissolving 10 mg MCLA in 34.5 mL MQ water, where upon 1 mL aliquots of this solution were then pipetted into 2 mL polyethylene vials and frozen at $-80^\circ$C until required for use. The working MCLA standard, 1 $\mu$M, was prepared from a thawed vial of the primary stock by dilution into a 1 L solution of 0.05 M Sodium acetate (Sigma Ultra) buffered (4.1 g) in MQ water adjusted to pH$_{\text{NBS}}$ 6. Several time points between 0 and 23 h were taken from the experimental solutions as described above, and were drawn directly into the flow cell of the chemiluminescence detector as described before (Heller and Croot, 2010c). The sensitivity of the MCLA method is very strongly temperature dependent and for this reason all samples and reagents were kept at constant temperature (21.5 $\pm$ 0.2$^\circ$C) throughout the course of the experiments.

**Analysis of $\text{O}_2^-$ Decay Using SOTS-1 as $\text{O}_2^-$ Source**

In this work the rate of the superoxide reaction with unamended seawater is reported as $k_{\text{SW}}$ while that of the DTPA amended is listed as $k_{\text{DTPA}}$. It is assumed that the rate of reaction with organic matter is equal to the reaction in the presence of DTPA, i.e., $k_{\text{org}} = k_{\text{DTPA}}$.

Other assumptions include:

(i) That the response of the system is overall first order with respect to $\text{O}_2^-$

(ii) That all the metal species are made inert by complexation with DTPA, leaving only reactions with organic species and the uncatalysed self dismutation reaction as the pathways for $\text{O}_2^-$ decay.

(iii) That the 2nd order self-dismutation reaction is significantly small that it can be ignored. In the present case this a reasonable assumption as this reaction is well-described in seawater as a function of pH (Zafiriou, 1990; Heller and Croot, 2010c) and the mM to pM levels of $\text{O}_2^-$ generated using $\mu$M concentrations of SOTS-1 indicates that results in a pseudo first order reaction on the order of $1 \times 10^{-4}$ s$^{-1}$ or less.

The analysis of the time dependent concentration of $\text{O}_2^-$ generated in seawater samples due to the additions of SOTS-1 was performed as previously described (Heller and Croot, 2010a). Briefly raw photon counts are firstly corrected for the signal blank due to MCLA auto-oxidation and then converted to a concentration of $\text{O}_2^-$ by applying the calculated sensitivity factor previously determined by calibration of a seawater sample with additions of KO$_2$.

The rate equation derived for the formation of $\text{O}_2^-$ from SOTS-1 previously (Heller and Croot, 2010a) is shown below:

$$\frac{d[\text{O}_2^-]}{dt} = 0.4k[SOTS]_0e^{-kt} - k_{\text{obs}}[\text{O}_2^-]$$

where $t$ is the time since the introduction of the SOTS-1, $k$ is the rate of thermal decomposition of SOTS-1, [SOTS]$_0$ is the initial
concentration of SOTS-1 and \( k_{obs} \) is the observed 1st order loss rate of \( O_2^- \). If the initial concentration of \( O_2^- \) is zero (a reasonable assumption when the seawater is filtered and kept in the dark), then Equation 1 has the exact solution (Harcourt and Esson, 1866):

\[
[O_2^-]_i = 0.4k[\text{SOTS}]_0 \left( \frac{e^{-kt} - e^{-k_{obs}t}}{k_{obs} - k} \right) \tag{2}
\]

When \( k_{obs} > k \), as would be expected under most circumstances in seawater, Equation 2 reduces to:

\[
0.4k[\text{SOTS}]_0 e^{-kt} = k_{obs} [O_2^-]_i \tag{3}
\]

Taking the natural logarithm of both sides then gives:

\[
-kt = \ln[O_2^-]_i + \ln \frac{k_{obs}}{0.4k[\text{SOTS}]_0} \tag{4}
\]

Thus, in this case a plot of \( \ln[O_2^-]_i \) vs. time will have a slope of \( k \), the thermal decomposition rate of SOTS-1. Note that the value of \( k_{obs} \) can also be determined here from the value of the intercept as the value for \( [\text{SOTS}]_0 \) is known. Rearranging Equation 3 gives the following:

\[
k_{obs} = \frac{0.4k[\text{SOTS}]_0 e^{-kt}}{[O_2^-]_i} \tag{5}
\]

Operationally the determination of \( k_{obs} \) is optimal once the maximum value of \( [O_2^-]_i \) is reached:

\[
t = \frac{\ln \left( \frac{k}{k_{obs}} \right)}{k - k_{obs}} \tag{6}
\]

Typically it takes 10 min to reach the maximum \( [O_2^-]_i \) at 25°C if \( k_{obs} = 0.01 \, \text{s}^{-1} \). Thus, in the present case we used data collected after 30 min and up to 23 h since the initiation of the experiment to determine \( k_{obs} \).

**RESULTS**

**Hydrography and Nutrients**

During the month of April, 2011 there was weak but persistent upwelling across the Mauritanian shelf with high concentrations of chlorophyll extending out into the Atlantic Ocean (Figure 2). Vertical profiles of chlorophyll fluorescence (Figure 3) indicated highest surface concentrations in offshore waters with lower levels inshore, indicating that satellite estimates of chlorophyll in these type 2 waters are currently overestimated, presumably by a combination of Saharan dust, suspended sediment, and CDOM. Transmission data (not shown) indicated that there was a considerable bottom nepheloid layer present across the shelf and shelf edge this was consistent with earlier microstructure measurements taken in the same region (Schafstall et al., 2010). The core of the OMZ of the ETNA was found in offshore waters between 300 and 600 m depth, as observed previously (Stramma et al., 2009) and was still present at this depth upon encountering the continental shelf. Shelf waters were still oxygenated however with a strong vertical gradient present between surface and depth.

**CDOM Absorbance**

Vertical profiles of CDOM absorbance, \( a_{325} \), at the 5 GO-FLO stations are shown in Figure 3. Highest values of \( a_{325} \) were found near the shelf break (Stn 14–15) and decreased offshore.

Values of \( E_2/E_6 \) (Figure 4) were ~6 at stations on the shelf and in surface waters off the shelf. At stations 8–9 there was a monotonic increase in \( E_2/E_6 \) with depth which may have been caused by strong mixing at depth across the shelf as evidenced by the transmission profiles (not shown). Thus, it is likely that material from the sediments with a higher \( E_2/E_6 \) (∼9) was being resuspended and mixed through the water column in the vicinity of the shelf edge (Stns 8–9 and 6–7). There is limited data for \( E_2/E_6 \) values of marine porewaters but what data there is suggests that values increase with sediment depth (Dang et al., 2014), though it is likely that this is dependent on the rate of carbon and \( O_2 \) supply. As an increase in \( E_2/E_6 \) is thought to reflect a shift toward lower molecular weight compounds this may also represent diffusion of such material from the shelf sediments.

**PARAFAC Analysis and FDOM_{H}**

PARAFAC analysis of our complete 3D EEM FDOM data set identified 3 independent components (Table 1; Table S3 and Figures 5, 6). The calculated excitation and emission spectra for each of the 3 components is shown in Figure S1. Component C1 was similar to a marine fulvic (M-peak), \( \lambda_{ex}/\lambda_{em} = 240/412 \) and 320/412. Component C2 had elements that were similar to both classical terrestrial humics (A and C-peaks), \( \lambda_{ex}/\lambda_{em} = 240/412 \) and 320/412. Lastly component C3 was similar to the Tryptophan-like T-peak, \( \lambda_{ex}/\lambda_{em} = 280/330 \). In the present case it was not clear if the data was of sufficient resolution for PARAFAC to separate out more components, other approaches which assume the presence of specific chromophores may be able to resolve this better but would be reliant on assumptions made regarding the presence of specific chromophores in solution. Direct measurements of humic fluorescence (not shown) or FDOM_{H} (\( \lambda_{ex}/\lambda_{em} = 320/420 \)) correlated [Spearman’s rho, calculated using corr function in Matlab (Mathworks)] most strongly with C1 (\( \rho = 0.59, n = 252, p < 0.001 \)) but also had weaker correlations with C2 (\( \rho = 0.21, n = 251, p < 0.001 \)) and C3 (\( r = 0.41, n = 251, p < 0.001 \)) (Tables 1 and S3, Figures 5, 6). This differed from our earlier work (Heller et al., 2013) where only one of the components identified by PARAFAC was strongly correlated to FDOM_{H}.

The vertical distribution of C1 and C2 at the Go-Flo stations are shown in Figure 5. Profiles of C1 were relatively constant throughout the water column (∼0.02 R.U.), the only noticeable feature is small positive excursions below the chlorophyll maxima (Figure 3) at the 3 stations offshore and suggests a relationship to zooplankton grazing or microbial remineralization of sinking particles. At stations near the shelf break, C1 was slightly elevated in surface waters, in contrast with open ocean profiles (Heller et al., 2013) of FDOM_{H} in open ocean tropical waters which typically exhibit a strong photobleaching effect with a minimum
in the surface. This suggests that the productive waters in the upwelling region were a stronger source of C1, FDOM$_{H_2O_2}$ to surface waters than photobleaching was a sink. At the offshore stations, and on the shelf, C2 was low and almost constant ($\sim$0.01 R.U.) throughout the water column with only a slight enrichment in surface waters close to the shelf break.

The distribution of the tryptophan-like component, C3 (Figure 6), also showed some similarities to C2 with local maximum below the chlorophyll maxima and suggestive also of links to zooplankton or microbial activity. Although comparison with NH$_4^+$ (Figure 6), as a tracer of zooplankton activity, reveals that NH$_4^+$ concentrations are higher in the mixed layer and not coincident with the C3 maxima. However, as NH$_4^+$ is also taken up by phytoplankton/bacteria this approach may be too simplistic. Dissolved O$_2$ concentrations were also weakly correlated with C2 ($\rho = 0.36, n = 251, p < 0.001$) and C3 ($\rho = 0.45, n = 251, p < 0.001$).

**H$_2$O$_2$ Distribution**

H$_2$O$_2$ concentrations (Figure 4) were as expected high in the sunlit surface waters with low background levels in the aphytic zone. Overall surface waters were low in H$_2$O$_2$ compared to other open ocean sites in the Tropical North Atlantic not impacted by the ITCZ (Croot et al., 2004), though they were similar to observations near Cape Verde (Heller and Croot, 2010b). Most likely the lower surface concentrations observed here is reflecting strong consumption of H$_2$O$_2$ by the phytoplankton growing in the upwelling waters. On the shelf edge (stations 6–7) there were slightly enhanced levels of H$_2$O$_2$ in the bottom waters possibly indicating a sedimentary source, coincident with elevated levels of C1 and C3 also seen at this station. As pore water measurements from sediments along the same transect, taken on the same expedition, indicated that surface sediments were always oxic and that only on the shelf itself did sediments accumulate H$_2$S at depth (Dale et al., 2014). Thus, it is unlikely that these sediments are a source of H$_2$O$_2$ from the oxidation of Fe(II) released from the sediments, and it is more likely resulting from bacterial respiration of particulate carbon that has sunk to the sediments from the overlying productive waters.

**O$_2$ Decay Rates Using SOTS-1 as Source**

Rates of O$_2^-$ decomposition (Table 2; Table S4) as determined using SOTS-1 as a thermal source for O$_2^-$ are shown in Figure 7 for seawater with DTPA ($k_{DTPA}$) and unamended ($k_{SW}$). Values
of $k_{\text{DTPA}}$ and $k_{\text{SW}}$ were comparable to most other recent studies in Tropical waters (Table 2). Over the entire transect the reaction between $O_2^-$ and DOM was a major pathway ranging from 28 to 80% of the overall loss rate ($k_{\text{DTPA}}/k_{\text{SW}}$) with an average of $58 \pm 13$ (1 sd) %. This is slightly more than what was observed further to the south of the present study area (Wuttig et al., 2013a) and in strong contrast to the Southern Ocean where the organic pathway was found to be only minor (Heller and Croft, 2010c). There was no significant correlation ($\rho < 0.1, n = 36$) between either $k_{\text{DTPA}}$ or $k_{\text{SW}}$ with the 3 components identified by PARAFAC or FDOM$_{\text{H}}$. A significant correlation was however found between $k_{\text{DTPA}}$ and $S_{350-400}$ ($\rho = 0.58, n = 36, p < 0.002$), with a weaker correlation between $k_{\text{SW}}$ and $S_{350-400}$ ($\rho = 0.44, n = 36, p < 0.02$).

### DISCUSSION

#### CDOM Absorbance Along the Transect across the Mauritanian Upwelling

CDOM Absorbance in EBUS

There have been a few other studies of CDOM in EBUS (Kudela et al., 2006; Day and Faloona, 2009) and the North Atlantic (Kitidis et al., 2006; Nelson et al., 2007; Nelson and Siegel, 2013). In the Californian system, CDOM absorbance at 325 nm was found to be lowest in recently upwelled waters (Day and Faloona, 2009) with the implication that it was derived from primary production and not from sediment resuspension. In a related study (Kudela et al., 2006), CDOM was found to exhibit seasonal patterns in the spectral slope ($S_{350-400}$) related to both the strength of the upwelling and to the input of rivers to this region.

Sources of CDOM to the Mauritanian Upwelling

The main source of CDOM to our study region appears to be from new production by phytoplankton as evidenced by increases in $a_{\text{CDOM}}$ (Figure 3) coincident with the euphotic zone. However, there were also a potential source from the sediment as seen in the $E_2:E_3$ data (Figure 4) at the shelf break, which the increase in $E_2:E_3$ suggested an input of lower molecular weight CDOM. A high $E_2:E_3$ source from pore waters or sediments has not been described before to our knowledge and without pore water measurements to constrain these values we hesitate to assign them to any particular biogeochemical process. The $E_2:E_3$ values found in the present study were similar to that found in surface marine waters off the Portuguese coast (Santos et al., 2014), where a significant negative correlation was found between $E_2:E_3$ and β-GlCase activity tentatively suggesting that decreases in $E_2:E_3$ are associated with increases in microbial activity. Photobleaching has been shown to increase $E_2:E_3$ (Helms et al., 2008) and while this process was clearly occurring in surface waters it was apparently not fast enough to change the phytoplankton or microbial signal observed.

A recent study in the waters of the equatorial upwelling region of the Atlantic (Andrew et al., 2013) showed several lines of evidence that the CDOM found in surface waters there was derived from terrestrial components. They suggested that changes in the optical properties of CDOM associated with increases in AOU (Yamashita and Tanoue, 2008, 2009; Swan et al., 2009; Nelson et al., 2010; Yamashita et al., 2010) could be associated with redox changes in existing terrestrial material. A recent work (Aparicio et al., 2015) made a direct test of the hypothesis put forward by Andrew et al. (2013), by incubating microbial communities with a suite of organic compounds (glucose and acetate) and with or without humic matter. They found that new FDOM$_{\text{H}}$ was produced particularly when humic precursors were added, thus in accordance with the hypothesis. The results of Aparicio et al. (2015) also supported the hypothesis of Jørgensen et al. (2014) in which the less labile the precursor material is, the more humic fluorescence is generated.

The hypothesis of Andrew et al. is partly based on evidence from ultra-resolution mass spectral data for the presence of terrestrial lignins in seawater (Opsahl and Benner, 1997; Kujawinski et al., 2009). However, recent data was published

### TABLE 1 | PARAFAC analysis and identification of fluorophores.

<table>
<thead>
<tr>
<th>Component ex/em</th>
<th>Fluorescence Characteristics ex/em</th>
<th>Description and probable source</th>
</tr>
</thead>
<tbody>
<tr>
<td>C1 240/412 and 320/412</td>
<td>290–310/370–410</td>
<td>Marine fulvic “M” peak (Coble, 1996)$^a$</td>
</tr>
<tr>
<td></td>
<td>340/420</td>
<td>UV/Visible humic-like (Wedborg et al., 2007)$^p$</td>
</tr>
<tr>
<td></td>
<td>315/418</td>
<td>Marine humic material “C2” (Murphy et al., 2010)$^b$</td>
</tr>
<tr>
<td></td>
<td>335/400</td>
<td>Marine humic “C4” (Heller et al., 2013)$^b$</td>
</tr>
<tr>
<td></td>
<td>&lt;270–370/470</td>
<td>Humic like “C2” (Catalá et al., 2016)$^b$, peak “M” (Catalá et al., 2015b)$^b$</td>
</tr>
<tr>
<td>C2 270/470</td>
<td>&lt;260/400–460</td>
<td>Terrestrial Humic Substance “A” peak (Coble, 1996)$^a$</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Humic like (Dubnick et al., 2010)$^b$</td>
</tr>
<tr>
<td></td>
<td></td>
<td>Humic like “C1” (Jørgensen et al., 2011)$^b$</td>
</tr>
<tr>
<td></td>
<td>250/475</td>
<td>Humic like “C2” (Heller et al., 2013)$^b$</td>
</tr>
<tr>
<td></td>
<td>320/400</td>
<td>Humic like “C1” (Catalá et al., 2016)$^b$, peak “AV” (Catalá et al., 2015b)</td>
</tr>
<tr>
<td>C3 280/330</td>
<td>275/340</td>
<td>Tryptophan-like peak “T” (Coble, 1996)$^a$</td>
</tr>
<tr>
<td></td>
<td>280/328</td>
<td>Amino acids “C6” (Murphy et al., 2010)$^b$</td>
</tr>
<tr>
<td></td>
<td>280/330</td>
<td>Tryptophan “C2” (Jørgensen et al., 2011)$^b$</td>
</tr>
<tr>
<td></td>
<td>280/320</td>
<td>Tryptophan “C3” (Heller et al., 2013)$^b$</td>
</tr>
<tr>
<td></td>
<td>290/340</td>
<td>Tryptophan-like “C3” (Catalá et al., 2016$b$, Tryptophan-like (Catalá et al., 2015b)</td>
</tr>
<tr>
<td>Not observed</td>
<td>275/305</td>
<td>Tyrosine-like peak “B” (Coble, 1996)$^a$</td>
</tr>
<tr>
<td></td>
<td>280/305</td>
<td>“BT” protein-like (Wedborg et al., 2007)$^b$</td>
</tr>
<tr>
<td></td>
<td>280/310</td>
<td>Tyrosine “C5” (Jørgensen et al., 2011)$^b$</td>
</tr>
<tr>
<td></td>
<td>270/310</td>
<td>Amino acid-like “C4” (Catalá et al., 2016)$^b$, Tyrosine-like (Catalá et al., 2015b)</td>
</tr>
</tbody>
</table>

$^a$PARAFAC analysis.

$^b$PARAFAC analysis.
that indicated that some of the lignin precursor compounds or breakdown products (including the Vanillyl, Syringyl, and Cinnamyl phenols) were exuded by diatom cultures under optimal growth and different metal stress conditions (Rico et al., 2013; López et al., 2015; Rubino, 2015). Now it could be argued that the diatoms, or bacteria associated with them, were simply degrading existing lignin structures in seawater, however it is well-established that many marine organisms contain or synthesize polyphenols (Vreeland et al., 1998; Rico et al., 2013; Gómez et al., 2016) so there is a strong possibility that some of this material is exuded or released by grazing zooplankton or viral lysis.

**Potential for ROS Production from CDOM**

Recently there have been a number of studies that have examined the relationship between CDOM properties and ROS production in natural waters (Dalrymple et al., 2010; Peterson et al., 2012; Zhang et al., 2012; Powers and Miller, 2014). Dalrymple et al. (2010) in a laboratory study with freshwater humic substances observed that quantum yields for $\text{O}_2$ ($\Phi_{\text{O}_2}$) increased with increasing $E_2$:E3 values, however quantum yields for $\text{H}_2\text{O}_2$ ($\Phi_{\text{H}_2\text{O}_2}$) decreased with increasing $E_2$:E3. These authors suggested that this inverse relationship between $\Phi_{\text{O}_2}$ and $\Phi_{\text{H}_2\text{O}_2}$ was due to competitive formation of the $\text{O}_2$ and $\text{H}_2\text{O}_2$ precursors. Sharpless and Blough (2014) further suggested that the inverse correlation of $E_2$:E3 with $\Phi_{\text{H}_2\text{O}_2}$ could be related to enhanced rates of CDOM radical formation due to higher levels of aromatic donors.

In the present study we were only able to measure a “snap shot” of the $\text{H}_2\text{O}_2$ concentration (Figure 4) and so have no production rates, though earlier work (M68-3) we performed in the same region over diel cycles suggested a strong gradient in $\text{H}_2\text{O}_2$ production from inshore to offshore (unpublished data). Near surface values of $E_2$:E3 (Figure 4) are relatively constant (~6) perhaps indicating new production dominated over photobleaching (Helms et al., 2008) at this time. It was likely then that $\text{H}_2\text{O}_2$ concentrations and inventories in near surface waters were determined more by loss rates than production, as an earlier study offshore of this region found higher $\text{H}_2\text{O}_2$ surface concentrations in low chlorophyll waters (Steigenberger and Croot, 2008), thus presumably the high phytoplankton and bacterial abundances in the near sure waters resulted in lower $\text{H}_2\text{O}_2$ due to the presence of cellular peroxidases.
TABLE 2 | Compilation of rates (with and without DTPA) for the decay of O$_2$ in seawater during this work and from other studies.

<table>
<thead>
<tr>
<th>Study</th>
<th>Location</th>
<th>Station</th>
<th>Depth</th>
<th>k$_{SW}$ (s$^{-1}$)</th>
<th>k$_{DTPA}$ (s$^{-1}$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>This study (Table S4)</td>
<td>Mauritanian shelf</td>
<td></td>
<td>0–1000</td>
<td>0.0003–0.0169</td>
<td>0.0006–0.0249</td>
</tr>
<tr>
<td>Wuttig et al., 2013a</td>
<td>ETNA</td>
<td>GoFlo 1 (CVOO)</td>
<td>19–600</td>
<td>0.025–0.043</td>
<td>0.010–0.030</td>
</tr>
<tr>
<td></td>
<td></td>
<td>GoFlo 2</td>
<td>46–391</td>
<td>0.021–0.086</td>
<td>0.016–0.38</td>
</tr>
<tr>
<td></td>
<td></td>
<td>GoFlo 3</td>
<td>20–400</td>
<td>0.029–0.065</td>
<td>0.016–0.089</td>
</tr>
<tr>
<td></td>
<td></td>
<td>GoFlo 4</td>
<td>20–400</td>
<td>0.015–0.048</td>
<td>0.011–0.028</td>
</tr>
<tr>
<td></td>
<td></td>
<td>GoFlo 5</td>
<td>20–400</td>
<td>0.013–0.031</td>
<td>0.006–0.017</td>
</tr>
<tr>
<td></td>
<td></td>
<td>GoFlo 6</td>
<td>20–400</td>
<td>0.009–0.083</td>
<td>0.011–0.051</td>
</tr>
<tr>
<td>Heller and Croot, 2010c</td>
<td>Southern ocean</td>
<td>230–6</td>
<td>25–1000</td>
<td>0.014–0.041</td>
<td>0.013</td>
</tr>
<tr>
<td></td>
<td></td>
<td>236–5</td>
<td>25–2800</td>
<td>0.008–0.037</td>
<td>0.013</td>
</tr>
<tr>
<td></td>
<td></td>
<td>249–3</td>
<td>25–1000</td>
<td>0.009–0.021</td>
<td>0.013</td>
</tr>
<tr>
<td>Heller and Croot, 2010b</td>
<td>ETNA</td>
<td>10 (CVOO)</td>
<td>10</td>
<td>0.023</td>
<td>0.013</td>
</tr>
<tr>
<td>Heller and Croot, 2011</td>
<td></td>
<td>8</td>
<td>5</td>
<td>0.008</td>
<td>0.013</td>
</tr>
<tr>
<td>Roe et al., 2016</td>
<td></td>
<td>Station aloha</td>
<td></td>
<td>0.003–0.014</td>
<td>0.014</td>
</tr>
<tr>
<td>Roe et al., 2016</td>
<td></td>
<td>California current</td>
<td></td>
<td>0.006–0.017</td>
<td></td>
</tr>
<tr>
<td>Rose et al., 2010</td>
<td></td>
<td>WQN157-184</td>
<td></td>
<td>0.08–0.31</td>
<td>0.004–0.012</td>
</tr>
<tr>
<td></td>
<td></td>
<td>TRICHO_1-3</td>
<td></td>
<td>0.07–0.43</td>
<td></td>
</tr>
<tr>
<td>Powers et al., 2015</td>
<td>Gulf of Alaska surface and deep</td>
<td></td>
<td></td>
<td>0.004–0.012</td>
<td></td>
</tr>
<tr>
<td>Hansard et al., 2011</td>
<td></td>
<td>GoA1-GoA4</td>
<td>10–50</td>
<td>0.0167</td>
<td>0.004–0.012</td>
</tr>
</tbody>
</table>

FIGURE 7 | Vertical distribution of O$_2$ decomposition rates $k_{DTPA}$ (green triangles) and $k_{SW}$ (blue circles) at 5 stations along the transect in the Mauritanian Upwelling.

FDOM Distribution: Sources and Sinks in the Mauritanian Upwelling

Recent Advances in Our Understanding of FDOM

Two recent contributions to this field have helped make significant advances, firstly in a recent review article, Sharpless and Blough (2014) showed that fluorescence spectra of natural water samples were not simply a superposition of individual chromophores and instead they suggested it could be explained using a physical model incorporating charge transfer interactions between electron donating and accepting chromophores within the CDOM. In this context then the identification of specific individual chromophores from a seawater sample is highly unlikely but broad conclusions about the functional groups might be gathered. In this regard another recent advance was provided by Wunsch et al. (2015) who compared quantum yields and fluorescence properties of chromophores identified as being potential components of CDOM and compared them to 3D EEM and PARAFAC data held in the OpenFluor database (Murphy et al., 2014).

PARAFAC Identification of FDOM Components

In the present work we identified 3 FDOM components (Table 1) that were related to the traditional M (C1), A and C (C2) and T (C3) peaks. Our PARAFAC results are similar to those found in other recent studies (Table 1). In particular, a series of paper published by Catalá et al. (2015a,b, 2016) examined the FDOM
data gathered during the Spanish Malaspina circumnavigation of the globe. This work builds on an earlier study by Jørgensen et al. (2011) that used data collected during the Danish Galathea circumnavigation. In the global data set acquired by Catalá et al. they identified 4 FDOM species by PARAFAC (denoted here using the subscript M and their original designation), 2 humic-like M1C1 related to peaks A and C and M2C2 associated with peak M) and 2 protein-like fluorophores (M3C3 Tryptophan −T, M4C4 Tyrosine −B). They were able to estimate an overall turnover time in the deep ocean, for the different peaks M1C1 (A/C) 435 ± 41 years and M2C2 (M) 610 ± 55 years and peak M3C3 (T) 379 ± 103 years (Catalá et al., 2015b). Modeling of their data indicated a higher production rate of M2C2 than M1C1 (almost double) as a function of AOU. Comparison with the current work suggests that M2C2 ~ C1 and M1C1 ~ C2, with M3C3 ~ C3 (Table 1).

Production/Decomposition of FDOM by Phytoplankton, Zooplankton, and Bacteria

Laboratory experiments with axenic phytoplankton indicates that many species can produce significant concentrations of FDOM during both growth and senescence phases (Chari et al., 2013; Fukuzaki et al., 2014). Measured EEMs of axenic cultures of the diatom *Ditylum brightwellii* (Fukuzaki et al., 2014) were found to produce peaks (λ<sub>ex</sub>/λ<sub>em</sub> = 350/450) similar to peak C, which had commonly been associated as a terrestrial humic. The raphidophyte *Heterosigma akashiwo* also produced FDOM (λ<sub>ex</sub>/λ<sub>em</sub> = 370/450–470) similar to peak C but slightly red shifted from the diatom (Fukuzaki et al., 2014). In a further laboratory study (Romera-Castillo et al., 2011) grew phytoplankton under axenic conditions and also found that they exuded humic-like substances (λ<sub>ex</sub>/λ<sub>em</sub> = 310/392—similar to C1 in the current study). When bacteria were grown in the phytoplankton exudates, the fluorescence from peak M decreased and new substances fluoresced (λ<sub>ex</sub>/λ<sub>em</sub> = 340/440—similar to peak C2 in our study). Peak T was seen to increase in the cultures during the exponential growth phase and decrease later.

Humic material is both produced (Shimotori et al., 2009, 2012; Zhang et al., 2015) and consumed (Bussmann, 1999; Rocker et al., 2012) by bacteria in the ocean. Laboratory studies have also shown that bacteria can simultaneously remove fluorescence associated with FDOM peak M that had been exuded by phytoplankton, while producing DOM (λ<sub>ex</sub>/λ<sub>em</sub> = 340/440) similar to FDOM peak C (Romera-Castillo et al., 2011). The generation of FDOM<sub>T</sub> by remineralization of particulate material resulting in the consumption of O<sub>2</sub> in intermediate and deep waters is presumed to occur by microbial activity (Yamashita and Tanoue, 2008) and results in strong correlations between FDOM<sub>T</sub> and AOU and species which are similarly regenerated by remineralization processes (e.g., NO<sub>3</sub> and PO<sub>4</sub><sup>3-</sup>). A strong correlation is found between FDOM<sub>M</sub> and AOU in most of the world’s ocean basins (Jørgensen et al., 2011; Catalá et al., 2016) with the exception of the North Atlantic (Jørgensen et al., 2011; Heller et al., 2013). In the equatorial Atlantic De La Fuente et al. (2014) found support for this relationship between AOU and FDOM<sub>M</sub> (defined in their study as λ<sub>ex</sub>/λ<sub>em</sub> = 340/440). In situ evidence for the formation of FDOM<sub>M</sub> comes from dark incubations in the eastern north Atlantic (Lønborg et al., 2015) results of dark incubation experiments where marine humic-like materials (λ<sub>ex</sub>/λ<sub>em</sub> = 320/410 similar to FDOM<sub>M</sub>) were produced as a by-product of microbial DOM degradation. The study of Lønborg et al. (2015) also revealed that the protein-like fluorescence (λ<sub>ex</sub>/λ<sub>em</sub> = 280/320) can be used as a proxy for the dynamics of the labile dissolved organic nitrogen (DON) pool, opening up the potential for looking using FDOM to look at aspects of the nitrogen cycle.

Copepods have been observed to exude a humic-like substance that fluoresces at peak M (C1; Urban-Rich et al., 2006), however in the present study the distribution of this fluorophore was relatively constant throughout the water column and showed no indication it was sourced from grazing (Figure 5). The composition of the phytoplankton prey is also important to the FDOM formed via grazing (Urban-Rich et al., 2004), with a release of humic like material with a diet of either diatoms or dinoflagellates in the exponential growth phase, but feeding on senescent cells lead to an increase in protein like FDOM. Interestingly Urban-Rich et al. (2006) found a shift to lower wavelength humic-like material may reflect a unique zooplankton signal. However, in the present case it appears that PARAFAC is unable to resolve such small differences in spectral signals in order to determine these as separate individual components.

Phytoplankton cells may also release organic matter after viral lysis and this “viral shunt” is suggested to be a major source of DOM in aquatic systems (Wommack and Colwell, 2000). The impact on FDOM resulting from viral lysis of the marine phytoplankton *Micromonas pusilla* was recently studied by Lønborg et al. (2013), who found that protein-like FDOM (λ<sub>ex</sub>/λ<sub>em</sub> = 280/320–Tryptophan like) and humic-like FDOM (λ<sub>ex</sub>/λ<sub>em</sub> = 320/410 similar to FDOM<sub>M</sub>) was elevated 4.1 and 2.8 times, respectively in infected cultures. This pioneering study demonstrates that viral lysis must also be considered in the production of FDOM in seawater.

In the context of the present study, and on the balance of the evidence from the literature, it appears that component C1 is most likely produced by phytoplankton, while C2 is derived from bacteria, with the possibility that elements of C1 are precursors for C2. However, this is a simplistic viewpoint as the relatively uniform vertical distribution of C1 at offshore stations may be interpreted as implying that the bulk of this signal is recalcitrant DOM with more labile material being formed in the high production zones close to the shelf break. For component C3 there are a number of potential sources and sinks with grazing the most likely source.

Photobleaching and Photoformation of FDOM

It has previously been noted that EEM Peak C is the most photo-labile component of FDOM (Helms et al., 2013) as expected as it is presumably lost by direct photochemical reactions as its excitation spectrum stretches into the near UV and thus will be exposed to sunlight in near surface waters. Similarly peak B has also been identified as being prone to photobleaching (Helms et al., 2013). No component similar to peak B was identified using PARAFAC in the present work.
species is also possible, as a humic like material was produced from the photo irradiation of the protein tyrosine (peak B; Berto et al., 2016). Nitrate and nitrite may also act as a photosensitizers in the photo-transformation of phenol containing compounds to form FDOM (Calza et al., 2012). A further potential reaction that may result in changes to FDOM is from brommination of organic matter (Méndez-Díaz et al., 2014) due to the reaction of CDOM with Br$_2^-$ formed by the reaction of Br$^-$ and OH radicals (Zafiriou et al., 1987).

**Determining if the Source of FDOM is Terrestrial or Marine**

The Fluorescence index (FI = $\lambda_{ex,470}/\lambda_{em,520}$ at $\lambda_{em,370}$; Hansen et al., 2016) has been used to distinguish between DOM derived from terrestrial or microbial sources. In the present case there is almost no fluvial input to this region (Cotrim Da Cunha et al., 2009), though terrestrial humic material may be associated with Saharan dust (Williams et al., 2007; Paris and Desboeufs, 2013) deposited in this region as has been previously suggested for the supply of organic matter to lakes in the Alps (Mladenov et al., 2011). In the present study we found values of FI = 2.6 ± 1.2, this is considerably higher than the typical range of reported values for FI of 1.2–1.8, however Hansen et al. (2016) found that leachates of the marine diatom Thalassiosira weissflogii had FI values up to 3.5 following biodegradation and lower values when this was combined with exposure to light. Earlier work at an oligotrophic site in the North Pacific (Station Aloha), far from any freshwater inputs had lower values for FI ranging from 1.55 in the surface to 1.72 at 3500 m (Helms et al., 2013). Thus, it appears that FI when applied in a marine setting is not a good indicator of terrestrial input but could instead be an indicator or phytoplankton derived FDOM.

In the present study, our PARAFAC analysis did not detect a distinct chromophore that aligned to peak C (C2 was aligned to A and C), which was initially regarded as being of terrestrial origin, though it has also been shown to form in seawater due to microbial action (Romera-Castillo et al., 2011). Another metric commonly used to assess terrestrial vs. marine FDOM is the ratio of Peak M to Peak C fluorescence (M:C), which should correlate with the relative source strength of marine-derived FDOM ($\lambda_{ex,310}/\lambda_{em,410}$) vs. terrestrial FDOM ($\lambda_{ex,345}/\lambda_{em,445}$; Helms et al., 2013). Across the Mauritanian shelf we found M:C = 1.03 ± 0.19. Higher values (mean 2.09) have been observed in coastal Mediterranean surface waters (Para et al., 2010). In the context of the hypothesis by Andrew et al. (2013) that all marine humic material may be terrestrial in origin it is clear that there are no distinctly terrestrial only FDOM signals, as a marine source can also be found, and that testing of this hypothesis requires the application of other analytical techniques.

**CDOM Promoted O$_2^-$ Decomposition Pathways**

**O$_2^-$ Decomposition Rates across the Mauritanian Upwelling**

Across our study area (Table S4), values of $k_{DTPA}$ varied from 0.0003 to 0.0169 s$^{-1}$ (mean 0.0074 ± 0.0038 s$^{-1}$) while $k_{SW}$ ranged from 0.0006 to 0.0241 s$^{-1}$ (mean 0.0122 ± 0.0054 s$^{-1}$). These values are similar to what has been observed in other recent studies of O$_2^-$ loss rates (see Table 2).

**Potential CDOM/FDOM Components as Sinks for O$_2^-$**

A number of organic species (e.g., quinones, thiols), have been suggested previously as potential DOM sinks for O$_2^-$ in seawater (Heller and Croot, 2010b). The correlation we found between $k_{DTPA}$ and $S_{550-400}$ ($\rho = 0.58$, $n = 36$, $p < 0.002$) is suggestive of a reaction between superoxide and aromatic moieties such as quinones, phenols/polyphehols, or humics which absorb over this range (Wünsch et al., 2015). Quinones have previously been identified as the most likely candidates as they react rapidly with O$_2^-$ and can be involved in a catalytic cycle with regeneration of the original reactant and production of O$_2$ and H$_2$O$_2$.

$$Q + O_2^- \rightarrow Q^*^- + O_2$$ \hspace{1cm} (R1)

$$QH_2 + O_2^- \rightarrow Q^*^- + H_2O_2$$ \hspace{1cm} (R2)

$$Q^*^- + Q^*^- \rightarrow 2H^+ \rightarrow QH_2 + Q$$ \hspace{1cm} (R3)

where Q is the quinone, QH$_2$ is the hydroquinone and Q$^*$ is the semiquinone radical (Eyer, 1991; Roginsky et al., 1999). The semiquinone radical however can also generate superoxide by reactions with oxygen (Meisel, 1975).

$$Q^*^- + O_2 \rightarrow Q + O_2^-$$ \hspace{1cm} (R4)

If no catalytic cycle is able to be established, then the reactant is consumed in a stoichiometric fashion and the reaction products will favor either O$_2$ or H$_2$O$_2$ depending on whether it is a reducing or oxidizing reaction. Identification of quinones as FDOM by 3D EEM is complicated as typically the hydroquinone has a high fluorescence quantum yield than the quinone itself, with many of quinones having no apparent fluorescence (Ma et al., 2010). Indeed it has been suggested that carbonyl compounds may play a more important role in the FDOM signal than quinones (Ma et al., 2010). So while quinones are likely present in seawater there is currently no data on what concentrations they may be present in.

It should be noted that the present work was conducted in a lab and in the absence of solar irradiation. During daylight in the ocean there are also reactants that could be photochemically generated, for instance oxidation of tryptophan (or other phenoxy species) to the semi oxidized Tryptophan radical occurs on reaction with Br$_2^-$, itself formed from OH reactions in seawater (Zafiriou et al., 1987). The tryptophan radical species reacts rapidly with O$_2^-$ to form tryptophan hydroperoxide (Ehrenshaft et al., 2015), however it can also react with O$_2$ to form O$_2^{-}$ and so may be both a source and sink for O$_2^{-}$. Similarly, tyrosine and protein tyrosyl radicals react rapidly with O$_2^{-}$ to form hydroperoxides (Möller et al., 2012; Das et al., 2014) and these reactions may be important under certain conditions (e.g., in the sea surface microlayer).

**Polyphenols as Superoxide Sinks in Seawater**

Previously it has been assumed that only source of polyphenols is from oxidation of terrestrial lignins (Opsahl and Benner, 1997,
1998; Krachler et al., 2012) and that they can form iron humic complexes that stabilize riverine iron in seawater (Krachler et al., 2010, 2012, 2015). Indeed, CDOM optical properties have been used to predict lignin concentration in seawater (Fichot et al., 2016). Gallic acid, along with other polyphenols from peat lands, has also been suggested as a possible ligand for transporting Fe to the sea (Wu et al., 2016). However, several lines of direct and indirect evidence now point to a seawater source for some of these compounds:

(1) Recent work has shown that polyphenols were released in response to metal stress by the diatom Phaeodactylum tricornutum (Rico et al., 2013; Santana-Casiano et al., 2014). Gallic acid was only detected in iron-enriched diatom cultures.

(2) An earlier estuarine study (Maie et al., 2007) using size exclusion chromatography found that the “T-peak,” usually assigned to “Tryptophan-like” substances (Coble, 1996), typically of high molecule weight (Yamashita and Tanoue, 2003), could be separated into a high molecule weight protein signal and a lower molecular weight polyphenol signal. More recently an investigation of riverine humics (Pagano et al., 2012) showed that the polyphenol, tannic acid, gave a peak $\lambda_{\text{ex}}/\lambda_{\text{em}} \sim 270/370$ nm, similar to the “T-peak.”

(3) Analysis of the OpenFluor database (Murphy et al., 2014; Wünsch et al., 2015) indicated a number of polyphenol compounds as being similar to the excitation/emission spectra identified by PARAFAC in 3D EEM spectra of natural samples.

In light of this, the T-peak appears to be related to the phenol content of Typtophan rather than the protein component. In the current work, PARAFAC component C3 was similar to the classical “T-peak,” however there was no apparent correlation between C3 and $k_{\text{DTPA}}$ ($r = -0.19$, $n = 36$). Though, it is recognized that there is a wide range of reaction rates with O$_2$ for the different phenolic compounds (Taubert et al., 2003) that may be present in this FDOM pool so it might be expected that there is very poor or no correlation at all between $k_{\text{DTPA}}$ and C3. As noted previously there is a difficulty in discriminating between similar phenol compounds using PARAFAC (Rosco et al., 2006).

Several of the polyphenols are known sinks for O$_2^-$ (Taubert et al., 2003; Terpinc and Abramović, 2010) and while there is currently significant interest in them as antioxidants there is surprisingly little agreement about their rate of reactivity with O$_2^-$ as most studies focused on inhibition experiments. An example of this is Ferulic acid (FA), a cinnamyl phenol, and an oxidation product of lignin, that are particularly abundant in grasses and many herbaceous tissues (Opsahl and Benner, 1998). FA reacts with O$_2^-$ (Toda et al., 1991; Nasr Bouzaïene et al., 2015), to form a radical species which is resonance stabilized and can form the dimer curcumin and other products (Graf, 1992). FA was also one of the phenolic compounds excised by diatoms in the study by Rico et al. (2013). Gallic acid (GA) reacts moderately fast with O$_2^-$ (Table S5; $k_{\text{GA}} = 5.4 \times 10^6$ M s$^{-1}$ at 25°C; Taubert et al., 2003).

Overall studies have shown that compounds with pyrogallol or catechol moieties are the most rapid superoxide scavengers, and the gallate moiety was found to be the minimal essential structure for maximal reaction rate constants with superoxide (Bors and Michel, 1999).

**Polyphenols As Sources of O$_2^-$ in Seawater**

That some polyphenols (e.g., tannin, pyrogallol, or gallic acid) produce H$_2$O$_2$ in weak alkaline solutions from reactions between with O$_2$ has been known for over 150 years (Schönbein, 1860). The reaction is thought to proceed via O$_2^-$ as an intermediate and both thermal and photochemical pathways of H$_2$O$_2$ formation have been observed (Clapp et al., 1990). The presence of such compounds could result in an additional O$_2^-$ thermal flux in experiments utilizing SOTS-1, though at typical seawater temperatures and assuming nM concentrations of these polyphenols this flux would be expected to be relatively insignificant.

**Consideration of Other Sources of FDOM to the Study Region**

In our study region there are no major riverine sources (Cotrim Da Cunha et al., 2009) to the coastal and thus a terrestrial/fluvial source for organic matter is considered unlikely. However, sandwiched between the upwelling region and the Saharan dust is the Banc d’Arguin, a shallow gulf with extensive tidal flats covered with extensive seagrass beds, predominantly Zostera noltii and Cymodocea nodosa (Hemmings and Nieuwenhuize, 1991). Many species of seagrasses have been found to contain high concentrations of polyphenols (e.g., Gallic, Caffeic, and Ferulic acid; Vergeer and Develi, 1997), including Zostera noltii (Grignon-Dubois et al., 2012) and Cymodocea nodosa (Cariello et al., 1979). Decaying seagrass (Hemmings and Nieuwenhuize, 1991) could thus be a potential source of poly phenols and/or FDOM to our study area, though from the limited physical oceanography carried out over the Banc d’Arguin the evidence suggest there is little exchange between the warm salty inshore water and the offshore upwelling (Peters, 1976; Loktionov, 1993; Carlier et al., 2015) and most of the seagrasses decay in situ (Hemmings and Nieuwenhuize, 1991).

**CDOM Red Fluorescence in OMZ Waters**

Previously Rötgers and Koch (2012) had reported the presence of a distinct absorption shoulder in CDOM at 415–420 nm in the OMZ waters of the Atlantic that was partially correlated with AOU. On further examination using methanol extracts of CDOM they related the absorption peak at $\sim$415 nm to red fluorescence at 650 nm. Rötgers and Koch suggested this peak may be bacterial in origin and most likely a non-chlorin, metal-free porphyrin, like degradation products of hemes, cytochromes, and chlorophyll c, etc. In an earlier work red fluorescence (Ex/Em 420/660) had been found to correlate with dissolved O$_2$ in the OMZ of the Arabian Sea (Breves and Reuter, 2000; Breves et al., 2003) with values ranging from 0 to 0.01 Raman units nm$^{-1}$ (Ex/Em 270/300 Heuermann et al., 1995). Conversely while a small absorption line at 420 nm was also reported for CDOM data from the Equatorial Atlantic (Andrew et al., 2013), those authors did not observe any related fluorescence at 640 nm and could not exclude it as a filtration artifact. In our work we observed fluorescence in the dissolved phase at these excitation/emission
wavelengths but there was no apparent correlation with oxygen concentrations (Figure 8) and the signal appeared to be related more to soluble pigments released in the photic zone presumably by zooplankton grazing (Kleppel, 1998).

AUTHORS CONTRIBUTIONS

MH, KW, and PC designed and planned the experiments. MH and KW participated in the cruise, carried out the sampling and analysis onboard. MH, KW, and PC performed the data analysis and wrote the paper.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: http://journal.frontiersin.org/article/10.3389/fmars.2016.00132

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